

WASTEWATER TREATMENT BY MICROALGAE

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abstract

Domestic wastewater contains high nutrients concentrations such as nitrogen and phosphorus which can affect public health and cause harmful ecological impacts. The potential of microalgae as a source of renewable energy based on wastewater has received increasing interest worldwide in recent decades. The microalgae cultivation in wastewater has two advantages: wastewater treatment and algal biomass production.

Our work aimed to remove nutrients from domestic wastewater by *Chlorella*.

Wastewaters were obtained from a wastewater treatment plant located in Ouargla, Algeria. The experiments were conducted in winter under natural sunlight in an outdoor open raceway pond situated in the desert area. Analysis of different parameters was done every 2 days along the period of the cultivation (19 days). The average removal efficiencies of COD, BOD₅, NH₄⁺-N and TP were maintained at 78%, 87%, 95% and 81% respectively.

Our results show the potential of integrating nutrients removal from wastewater by microalgae as a secondary wastewater treatment processes.

Key words : Microalgae, Wastewater, Nutrient removal, Secondary treatment, Desert.

1. INTRODUCTION

Large-scale microalgal production is limited by various factors that would otherwise ensure an economically feasible process.

These include selection of a cultivation system (open or closed system), utilizing efficient microalgal strains, the quantity and quality of light, the availability of nutrients and carbon dioxide (CO₂) to the algae and a reliable source of water that has little or no environmental impact (Lam and Lee, 2012).

Nevertheless, production processes are still under development and there is considerable scope to reduce costs and improve efficiency with this technology.

Microalgae cultivation requires large amounts of water (Rawat et al., 2013). The impact of large-scale algae production on water utilization has generated great debate.

Using seawater, brackish water or WW would reduce the need for freshwater (Komolafe et al., 2014).

Microalgae have been previously used for the treatment of WW (Komolafe et al., 2014; Ma et al., 2014). In turn, WW has been shown to be economically viable and sustainable for the production of microalgal biomass (Yang et al., 2015). Recent findings have shown that microalgae are able to grow and utilize the N and P

present in WW, thereby, removing these nutrients before WW discharge (Ma et al., 2014). Ammonia-N and P in secondary-treated WW are generally in the range of 20–40 mg/L and 10 mg/L, respectively (Grady et al., 2011). These nutrient concentrations are deemed sufficient to support the growth of most freshwater microalgae (Olguín, 2012). Coupling WW

treatment with biomass production is a very attractive option for energy, freshwater and fertilizer reduction. This would reduce the cost of WW treatment incurred for nutrient removal by conventional methods (Lam and Lee, 2012). Cultivation of microalgae in WW has the potential to reduce the N requirement

by up to 97% (Olguín, 2012). Wastewater utilization as a replacement for freshwater cannot totally negate the need for additional potassium, magnesium and sulphur (Rawat et al., 2013).

This work assessed the suitability of domestic WW as a medium for cultivating *Chlorella pyrenoidosa*. The experiments were conducted in the desert area in open pond systems. The study also aimed to evaluate and compare the nutrient removal efficiencies of algal secondary treatment and lagoon processing using bacteria for nutrient removal.

2. Materials and methods

2.1. Experimental site and wastewater

The study was carried out at a domestic WW treatment plant situated in the northeast of the Capital of Ouargla, Algeria (geographic coordinates: latitude: 31°59'46, 23 ' N; Longitude: 5°21'55, 77' E). The experiments were carried out in the winter period using sunlight as a light source. The temperatures ranged between 18 and 31 °C during the day and 6–15 °C at night. The duration of insolation during this season was approximately 9 h a day (7 a.m._5 p.m.); the maximum irradiance was approximately 1100 W/m². This was measured using a solarimeter (SL200, KIMO instrument, France). The WW used in this study was obtained directly from the aerated lagoon station basins after primary screening. The characteristics of the raw WW (RW) are summarized in Table 1.

Table 1. Wastewater characteristic before/after algae and lagoon process treatment

Characteristics	Raw WW	Algae treated WW	Lagoon treated WW
pH	7.82	9.36	7.43
Dissolved oxygen (mg/l)	0.2	9.4	4.5
Conductivity (mS/cm)	7.97	10.88	10.99
Salinity (g/l)	4.4	6.1	6.8
COD (mg/l)	426	90	110.75
NH ₄ ⁺ -N (mg/l)	46.2	2.1	32
TP (mg/l)	3.22	0.596	2.41
NO ₃ ⁻ -N (mg/l)	1.156	16.9	0.57

3. Results and discussion

3.1. Algal growth in wastewater

From the WW characteristics (Table 1), it can be seen that the amount of nutrient elements in WW is more than that of the prepared BG11 medium. The WW contains abundant N, available as $\text{NO}_3\text{-N}$ (1.15 mg/L) and $\text{NH}_4\text{-N}$ (46.2 mg/L) and P as TP (3.22 mg/L). The variation of temperature during the experiment ranged between 18 and 31 °C (Fig. 1). Zhao et al. (2015) showed that at a temperature of 35 °C, *C. pyrenoidosa* was able to achieve a biomass productivity of 0.16 g/L.d. From Fig. 1 it can be seen that the maximum temperature during the culture period reached as high as 31 °C. This can be considered low when compared to other seasons. During the summer (July–August) the temperature can reach 50 °C. The latitudes and the absence of desert area let the annual levels of solar radiation significantly higher than in other areas. High and uniform incident irradiance levels present an advantage for microalgal cultivation by increasing the photosynthetic active radiation (PAR) (Garcia-González et al., 2003), which represents around 43% of the solar radiation.

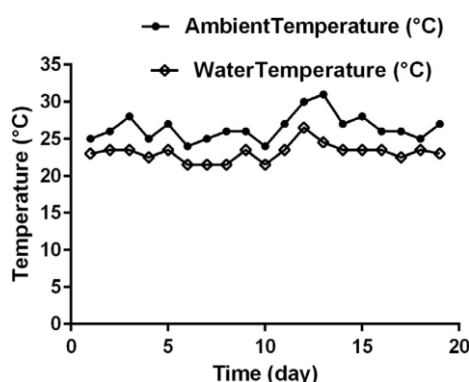


Fig. 1. Ambient and water temperature during the culture period

At the start of the experimentation, the pH was 7.8 (Fig. 2). This gradually increased to 9.2. This was mostly related to photosynthetic activity. As microalgae use CO_2 , the pH increases from acidic to alkaline due to the CO_2 depletion. The pH optimum for most algae range between 7 and 12. Fig. 2 shows the variation of the DO during the experimental period. It can be seen that the lowest DO was on day 1. This was potentially attributed to organisms such as bacteria that required the DO to decompose any organic material in the WW.

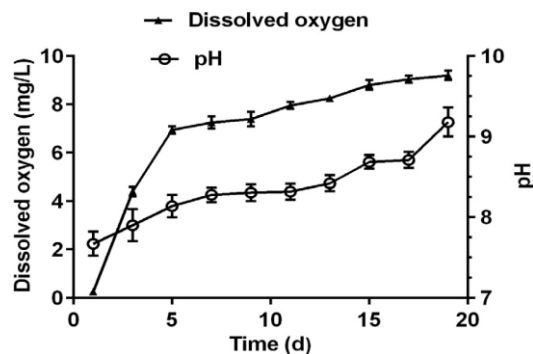


Fig. 2. Dissolved oxygen and pH variation during the culture period.

It then increased gradually to an approximate concentration of 8 mg/L during the cultivation period. From the results, it can be seen that the microalgal growth increases with DO. Microalgae perform photosynthesis in which they release DO into the culture medium. As the culture density increased, there was an increase in the DO levels in the medium.

4. conclusion

The *Chlorella* strain tested was able to grow satisfactorily in pretreated wastewater. The experimental results indicated that *Chlorella* has potential to remove COD concentration, ammonium and total phosphorus at a reasonable uptake rate from wastewater while being cultivated using wastewater as a culture medium.

The nutrient removal efficiencies were 78% (COD), 95% NH_4^+ -N and 81% (TP). The results in this study showed that *Chlorella* cultivation in wastewater is a promising method to produce algal biomass in addition to nutrients removal.

5. References

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